•综述•

土壤动物群落空间格局和构建机制研究进展

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摘要: 群落空间格局和构建机制一直是生态学研究的核心内容。在生物多样性严重丧失的背景下, 揭示群落空间格局及其构建机制, 有助于深刻理解生物多样性丧失的原因, 更有助于应对生物多样性保护等重大生态环境问题。然而, 陆地生态系统的研究多集中于地上生物群落, 对地下生态系统, 尤其是土壤动物空间格局和构建机制的研究尚不充分。事实上, 土壤动物多样性是全球生物多样性的关键组成之一, 是地下生态系统结构和功能维持的重要部分。对土壤动物空间格局和构建机制的研究, 能明确不同空间尺度条件下土壤动物多样性的维持机制。土壤动物群落常在多种空间尺度形成复杂的空间分布格局, 因此, 本文首先介绍了不同空间尺度主要土壤动物群落的空间自相关性特征, 阐述了土壤动物群落斑块和孔隙镶嵌分布的复杂空间格局。继而阐明这种空间格局主要受生物间作用、环境过滤和随机扩散的调控, 并说明这三个过程对土壤动物群落的调控能力和作用方式。作者提出, 这三个过程仍是今后土壤动物群落空间格局和构建机制研究的重点内容, 需要进一步加强以土壤动物为研究对象的群落构建理论的验证和发展。我国土壤动物群落空间格局和构建机制起步较晚, 希望本文能够促进我国土壤动物生态学相关领域的研究。

关键词: 空间尺度; 空间异质性; 多样性维持; 生物间作用; 环境过滤; 随机扩散

Spatial patterns and assembly rules in soil fauna communities: A review

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Abstract: Identifying spatial patterns and assembly rules in communities is a central study topic in ecology. With the unprecedented rate at which biodiversity is decreasing, it is necessary to recognize the spatial patterns and assembly rules in communities in order to understand why biodiversity is being lost and to be able to protect it. However, previous studies have focused more on plant communities in above-ground terrestrial ecosystems, neglecting below-ground ecosystems, especially soil faunal communities. Indeed, soil faunal biodiversity is a crucial component of global biodiversity because soil faunal communities assist in the maintenance of important ecosystem structures and functions. Therefore, one important aim of identifying spatial patterns and assembly rules in soil faunal communities is to clarify mechanisms of maintaining soil faunal biodiversity at multiple scales, so as to promote these processes, which also maintain ecosystem structures and functions. Soil faunal communities usually form complicated spatial patterns at multiple spatial scales. Here, we propose spatial autocorrelation characteristics, and then show how the complicated spatial patterns are demonstrated by patches and gaps of soil faunal communities at multiple scales. These spatial patterns are mainly controlled by processes of biotic interactions, environmental filtering and random

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dispersal. Consequently, we discuss the impacts of these processes on soil faunal communities. Finally, we suggest that these three processes are essential to evaluate and construct a theoretical framework for soil faunal communities and should continue to be studied in the future. Because interest in spatial patterns and assembly rules of soil faunal communities is relatively new in China, we expect this review will promote the development of related research areas.

Key words: spatial scale; spatial heterogeneity; biodiversity maintainance; biotic interaction; environmental filtering; random dispersal

群落格局和构建机制(维持机制)一直是群落生态学和生物地理学的核心问题(Huston, 1994; 牛克昌等, 2009), 其中最关键的是物种共存机制, 生态位理论和中性理论围绕这一问题展开了激烈的争论。生态位理论强调确定性因素是主导过程, 认为群落物种共存以生态位分化为前提, 有相同生态位的物种因为对共同资源的喜好产生激烈竞争, 这种竞争排斥作用致使物种不能稳定共存(Diamond, 1975)。然而传统的生态位理论在解释婆罗洲热带雨林物种多样性时遇到了挑战(Bell, 2000; Hubbell, 2001), 相关研究认为该生境内无明显限制性资源、无明显多物种生态位分化, 群落内物种却依然稳定共存。在此背景下中性理论开始备受关注, 该理论认为相似物种可以共存, 强调扩散和随机作用是群落构建的主导过程(Hubbell, 2001)。然而生态位理论和中性理论并非对立, 二者共同解释土壤动物群落的构建, 只是在不同空间尺度下其作用强度不同(Gao et al, 2014)。研究表明土壤动物群落主要受三个过程的调控, 即扩散(dispersal)、环境过滤(environmental filtering)和生物间作用(biotic interaction), 它们在多尺度上调控群落物种共存和维持生物多样性(Leibold et al, 2004; Emerson & Gillespie, 2008)。量化这三个过程的相对作用强度, 是确定生态位理论和中性理论相对贡献的基本途径, 也是群落生态学揭示地下群落构建机制的主要途径(Bell, 2010; Caruso et al, 2011)。

土壤动物物种多样、数量丰富、分布广泛, 参与物质分解、营养供应等生态学过程并为人类提供关键的生态系统服务(Ettema & Wardle, 2002; Lavelle et al, 2006)。尽管土壤动物多样性是全球生物多样性最关键的组成之一, 但目前相关研究主要集中于地上植被和动物群落(Ettema & Wardle, 2002; Whittaker & Heegaard, 2003), 对土壤动物群落空间格局和构建机制研究明显不足(Bardgett, 2002), 这

阻碍了地下生态系统多样性维持机制研究。本文首先综述了土壤动物群落多尺度空间格局, 通过阐明生物间作用、环境过滤和随机扩散对土壤动物群落构建的相对重要性, 说明生态位理论和中性理论对土壤动物群落维持的相对解释能力, 总结和展望其发展趋势和研究重点。通过综述土壤动物群落空间格局和构建机制这一群落生态学的核心问题(Heike et al, 2009; Lindo & Winchester, 2009; Decaëns, 2010; Berg, 2012; Caruso et al, 2012), 以期促进我国土壤动物群落空间格局和构建机制研究。

1 土壤动物群落空间格局和构建机制的尺度效应

土壤动物群落空间格局和构建机制具尺度效应(Borcard & Legendre, 2002; Ettema & Wardle, 2002), 即在不同的空间尺度条件下其群落格局和构建机制不同(Chase & Leibold, 2002), 且一个群落可能同时受到来自多个尺度过程的调控(Levin, 1992), 所以量化空间尺度是理解群落空间格局和构建机制的关键(Cadotte & Fukami, 2005; Gascoigne et al, 2005)。空间尺度参数包括空间幅度、空间粒度和聚焦尺度(Lundholm, 2009)。空间幅度是样方所在的全部面积大小; 空间粒度是异质性测量的尺度, 即斑块大小, 一般认为是邻近测量样点的平均空间距离; 聚焦尺度是测量物种多样性的尺度, 例如采集土壤动物的样方大小。其中空间幅度受到了较多关注, 尽管空间幅度从研究样地到整个大陆是连续变化的, 但为了研究方便常划分为离散尺度, 例如小尺度、局地尺度、区域尺度和全球尺度(Leibold et al, 2004)。在群落物种构建机制研究中, 空间粒度比其他尺度参数更重要, 因为物种是对环境距离斑块响应而不是对抽象的异质性指标响应(Tamme et al, 2010)。引起异质性的原因也可能因尺度而异(Ettema & Wardle, 2002), 在不同的空间尺度

条件下,构建机制对群落的调控作用又因环境条件而异(Bardgett et al, 2005),这使群落格局和构建机制的研究更加复杂。本文主要关注不同空间幅度内土壤动物群落空间格局和构建机制,分别关注小尺度、局域尺度和区域尺度的研究现状。

2 土壤动物群落空间格局

由于计算机技术的发展,空间直观方法(spatially explicit methods)在生态学中的应用越来越广泛(DeAngelis & Yurek, 2017),该方法结合生态学和空间数据,可以在空间维度直观地展示、评估、模拟和预测生态学格局和过程(Sala et al, 2010),尤其是那些具有空间依赖性的生态学过程(DeAngelis & Yurek, 2017)。基于空间直观方法有助于揭示土壤动物群落多尺度空间格局(Beale et al, 2010; Blanchet et al, 2013; Jiménez et al, 2014)和群落构建机制理论,包括源-汇理论(Source-sink Theory)、物种共存的空间贮存效应(Storage Effect)、生物多样性的中性理论、物种丰富度的竞争-拓殖权衡假说(Competition-colonization Trade-off) (DeAngelis & Yurek, 2017)等。本文着重介绍基于空间直观方法的土壤动物群落水平空间格局研究进展。基于空间直观方法的土壤动物格局研究开展较晚(Ettema & Wardle, 2002)。尽管研究表明空间格局可以在多种尺度进行预测,但是土壤动物群落的异质性常常被认为是“噪音”,使人难以辨别土壤动物群落驱动哪些生态过程,以及哪些变量和过程调控土壤动物群落(Smith & Wallace, 1976)。因此,长期以来空间异质性被认为是理解土壤动物群落构建机制和阐明其生态系统功能的阻碍。现在则普遍认为空间异质性不仅不是障碍,反而是理解土壤动物群落格局和构建机制的关键(Smith & Wallace, 1976)。

2.1 空间自相关性

空间自相关性被称为地理学第一定律(Tobler, 1970),即空间邻近的群落比空间远离的群落有更大的相似性。空间自相关性是生态学和地理学数据最常见的现象(Legendre, 1993; Kissling & Carl, 2008),该现象的存在意味着土壤动物群落分布具有内在的结构性,这是土壤动物群落特定空间格局形成的重要基础。大部分土壤动物群落具有显著的空间自相关性(Gutiérrez-López et al, 2010),在小于

100 m 的尺度内很少有空间独立性(Robertson, 1987)。土壤动物群落的空间自相关性依类群和生境而异,如蚯蚓、跳虫、螨类和甲虫等主要土壤动物,在不同生态系统均具有不同的空间自相关性特征(表1)。不同土壤动物类群体型大小各异,其空间自相关性距离也不尽相同:通常体型较大的土壤动物会出现较大的空间自相关性距离(Ettema et al, 2000; Decaëns & Rossi, 2001)。但尚没有研究明确提出不同体型土壤动物具有明显不同的空间自相关性特征,因为除了体型大小之外,土壤动物的空间自相关性距离还取决于具体的实验设计,例如空间幅度、空间粒度。如研究发现马德里附近一个森林生态系统中,蚯蚓、跳虫和螨类具有不同的空间自相关性距离,但这些不同体型土壤动物的空间自相关距离并没有显著差异(Gutiérrez-López et al, 2010)。

表1 主要土壤动物群落的空间自相关性距离

Table 1 Spatial autocorrelation distances of main soil animal communities

群落 Community	生境 Habitat	空间幅度 Spatial extent (m ²)	空间自相 关距离 Range (A ₀ , m)	参考文献 Reference
土壤蚯蚓 Earthworm	草地 Grassland	—	1–50	Decaëns et al, 2009
	森林 Forest	10,000	17–50	Gutiérrez-López et al, 2010
土壤跳虫 Soil collembola	森林 Forest	10,000	20–27	Gutiérrez-López et al, 2010
	农田 Cropland	250	5–50	高梅香等, 2014c
土壤螨类 Soil mite	森林 Forest	25	0.5–5	高梅香等, 2014a
		10,000	17–30	Gutiérrez-López et al, 2010
	农田 Cropland	250	5–40	高梅香等, 2014b
土壤线虫 Soil nematode	草地 Grassland	1,500	4–60	Rossi & Quénéhervé, 1998
	农田 Cropland	1,200	9–91	Liang et al, 2005
地表甲虫 Ground beetle	森林 Forest	25	1–2.5	胡媛媛等, 2018
		400	5–10	朱纪元等, 2017
		900	150–450	倪娟平等, 2018b
	农田 Cropland	1,600	40–150	刘洁等, 2017

因为较大空间尺度的研究需要大量的空间采样点数据, 因此目前土壤动物群落空间自相关性研究多集中于较小的空间幅度。但这并不说明土壤动物群落在较大空间幅度没有空间自相关性, 相关问题还有待进一步探讨。

土壤动物群落的空间自相关性可以由内因性过程引起(即固有空间自相关性) (Fortin & Dale, 2005), 包括与距离相关的繁殖、扩散、物种形成、灭绝和地理范围扩展(Legendre, 1993; Diniz et al, 2003); 也可以由外因性过程引起(即诱导空间依赖性) (Fortin & Dale, 2005), 包括地貌过程、风、能量输入和气候限制等空间结构环境因子。空间邻近的群落一般构建于更相似的环境条件, 这些相似的环境因子能够导致物种种群形成某种空间结构(Legendre, 1993; Diniz et al, 2003)。例如Lamto热带生态站(Station d'Ecologie Tropicale de Lamto)的2个蚯蚓种群(*Chuniodrilus zielae*和*Sthulmania porifera*) (Rossi, 2003a)的空间自相关性很相似, 它们的空间分布至少部分取决于相同的生态因子, 或者由那些在多种空间尺度起作用的不同生态因子调控。但即使在微空间尺度, 两个空间邻近但环境并不相似的地点, 其群落结构也会不同(Caruso et al, 2012)。

2.2 空间格局

土壤动物常形成高密度(即斑块)和低密度(即孔隙)镶嵌分布的异质性空间格局(Perry et al, 1999), 这里的斑块定义为局地物种密度显著高于平均密度, 孔隙定义为局地物种密度显著低于平均密度(Rossi, 2003a)。在小空间尺度, 土壤动物群落的特定物种关系会形成斑块性空间格局, 这些空间格局的空间范围取决于土壤动物的体型大小、生态系统类型和研究采样的空间设计(Ettema & Wardle, 2002)。尽管这些空间格局有镶嵌性和复杂性, 当实验数据充足的时候也可能发现一些普遍性特征。例如, 不同类群的土壤动物常形成不同大小的集群, 线虫(Ettema et al, 2000)和线蚓(Lavelle & Spain, 2001)常形成几厘米的集群, 蚯蚓常形成几米、几十米的集群(Decaens & Rossi, 2001; Margerie et al, 2001; Rossi, 2003b; Decaëns et al, 2009)。而蚂蚁巢穴的空间集群并不明显, 形成过度分散的格局(Traniello & Levings, 1986)。

群落内单个物种种群也可以形成聚集型空间分布格局。如两种内栖蚯蚓(*Glossodrilus* sp. 和

Andiodrilum sp.)随时间呈现出稳定的集群格局, 该格局可能由这些物种的种群繁殖率和生活史性状引起(Jiménez et al, 2006)。这些蚯蚓的高种群繁殖率使它们快速增长, 它们的内栖行为和中等体型意味着较低的扩散能力, 因此两个物种有可能呈现出显著的聚集型分布(Jiménez et al, 2006)。体型较小的蚯蚓(*Aymara* sp.和*Ocnerodrilidae* sp.)也呈现出聚集型的空间分布格局, 这些物种多聚集于有机质含量高的区域, 如牛粪、粪堆和甲虫巢穴(Jiménez et al, 1998; Decaëns, 2000), 这些物种的空间分布反映了营养资源可获得性对空间格局形成的重要性(Jiménez et al, 2006)。

物种-面积关系是土壤动物群落空间格局的重要基础, 对于一个给定的类群, 其物种丰富度随空间尺度的增加而增加(MacArthur & Wilson, 1967; Gaston, 2000), 调控该物种-面积关系的机制随尺度明显变化。在小尺度空间, 随着面积的增加, 调查到稀有物种的可能性随之增大, 结果可以调查到更多的物种。在较大的空间尺度, 增加的面积可能导致高的生物多样性和环境稳定性, 因此显示较高的迁入和局地灭亡比例。土壤动物群落的物种-面积关系报道较少, 见于马来群岛的蚂蚁(Wilson, 1961)、岛屿和大陆的螨类(Maraun et al, 2007)、跳虫(Ulrich & Fiera, 2009)和欧洲蚯蚓等(Judas, 1988)。如在一个给定的面积内, 岛屿上蚂蚁的物种丰富度少于大陆(Wilson, 1961), 这显示了生境隔离对群落物种丰富度的负效应。

我国土壤动物群落空间格局研究主要在森林和农田开展, 空间幅度涉及5、20、40、50、300和400 m, 空间粒度涉及0.5、2、5、20、40和50 m, 分别以土壤线虫(Liang et al, 2005; 梁文举等, 2006; Hua et al, 2008)、跳虫(高梅香等, 2014c)、螨类(高梅香等, 2014a, b)、步甲、隐翅虫、葬甲(朱纪元等, 2017; 胡媛媛等, 2018)和其他地表大型节肢动物(刘洁等, 2017)为对象, 开展了多尺度群落空间格局研究(图1)。研究表明, 空间自相关性是土壤动物群落的普遍现象, 土壤动物群落通常在小于150 m的距离内呈现显著的空间自相关。土壤动物群落的空间格局可以用指数、球状、高斯和线性模型进行拟合, 其空间格局多表现为斑块和孔隙镶嵌分布的复杂格局, 这些格局受地形、土壤或植被因子的影响(Hua et al, 2008; 高梅香等, 2015, 2018)。这些空间格局特

图1 凉水9 ha (300 m × 300 m)阔叶红松林样地地表步行虫群落(2015年8月)和地下土壤跳虫群落(2014年9月)克里格空间插值格局图(高梅香等, 2018)

Fig. 1 Kriging prediction maps for ground carabid (August 2015) and soil collembolan (September 2014) communities (Gao et al, 2018)

征依空间尺度和研究类群而异, 且在群落和种群水平表现规律不同, 通常优势物种和常见物种种群的空间格局特征更明显。

3 土壤动物群落构建机制

揭示群落构建机制的一个重要途径就是阐明调控群落格局的典型过程, 这主要包括扩散、环境异质性和生物间作用。扩散指的是物种沿着生境或景观从一个地点移动到另一个地点, 扩散决定到达某个地点的物种集合(Logue et al, 2011; Baguette et al, 2013)。环境异质性是一个复杂的概念, 指与环境相关的空间复杂性、多样性、异质性和结构性的综合(Stein et al, 2014), 环境异质性决定一个局地群落是否包含或排除某些物种, 因为它们的功能性状可能匹配或不匹配局地环境特征。生物间作用既包括种内作用(intraspecific interaction), 也包括种间作用(interspecific interaction), 如捕食、寄生、竞争等, 可以决定物种间的时空分布并产生多种群落结构(Gascón et al, 2016)。扩散、环境异质性和生物间作用交互影响, 被认为是一系列的过滤器, 共同决定群落格局, 因此环境异质性的调控作用也被称为环境过滤。这三个过程的调控能力依尺度而异: 通常认为在较大的区域尺度, 受气候条件和地理障碍影响的扩散过程决定群落格局, 并决定区域物种库对局域群落物种组成的影响(Dong et al, 2016); 在局域尺度, 环境过滤可能决定个体的迁入和迁出(Dong et al, 2016); 而在小尺度, 生物间作用则更加重要(Gascón et al, 2016) (图2)。需要注意的是, 土壤动物群落格局可能同时受到这三个过程的调控, 只

图2 群落构建机制的过滤模型(仿Vellend, 2016)。在这个假设的案例中, 一个来自区域物种库的随机物种子集进入局地地点; 只有那些有圆形边缘的物种能够适应局地环境条件; 竞争淘汰了其他物种, 只留下每个功能类型(以不同形状表示)的一个物种。由于资源需求不同, 两个剩余物种才能稳定共存。

Fig. 2 The filter model of community assembly (imitating from Vellend, 2016). In this hypothetical example, a random subset of species in the regional pool has access to a local site; only species with rounded edges can tolerate the environmental conditions; and competition leads to the elimination of all but one species of each functional type (shape). The two remaining species can coexist stably given contrasting resource requirements.

是在不同空间尺度、对不同土壤动物类群的调控强度不同(Vellend, 2016)。另外, 任何一种过程对群落的调控都是依群落特定情况而异(Vellend, 2016), 如移除捕食者有时会引起群落格局变化, 有时却对群落格局并无影响(Shurin et al, 2002)。因此, 研究不同空间尺度扩散、环境过滤和生物间作用对土壤动物群落的相对贡献, 对揭示土壤动物群落构建机制更为重要(Gascón et al, 2016)。

3.1 生物间作用

土壤动物多样性受相同营养级内生物间作用的潜在调控, 尤其是那些受资源影响而不受消费者影响的群落, 因为这些群落的结构最有可能受到竞争的调控(de Ruiter et al, 1995; Wardle, 2002)。种内竞争影响个体表型性状, 例如体型大小和生长速率(Anderson & Whiteman, 2015), 这些性状是体现物种适合度的重要基础(Kingsolver & Pfennig, 2004; Anderson & Whiteman, 2015), 进而可能影响群落水平的一些属性或特征。例如, 种内竞争导致不同大小越冬幼虫的生长能力差异, 而这种生长能力的差异可能会强烈地影响群落结构和组成(Anderson et al, 2013; Anderson & Whiteman, 2015)。但种内竞争的调控作用在土壤动物群落研究中相对少见, 如种内竞争调控食物和其他资源, 进而影响蚂蚁巢穴的空间格局(Traniello & Levings, 1986)。

局地尺度内相同营养级的种间竞争对土壤动物群落构建也起重要作用。Diamond (1975)指出种间竞争会妨碍那些生态学特性相似的物种在群落内共存。生态位重叠度用来衡量两个或多个物种对相同资源的利用, 那些在任何生态位维度都过度重叠的物种不能在相同斑块内共存, 而那些在相同斑块内共存的物种必须在某些资源利用或生态位维度上(如体型大小或营养级形态)有所分化, 才能利用不同的资源。在哥伦比亚东部平原的一个热带牧场, 相似的蚯蚓物种在不同的隔离斑块内呈现空间隔离分布, 尽管这些物种在垂直分布和体型大小方面有微弱差异, 但是在同一斑块内它们的形态过于相似, 对资源有强烈的潜在竞争(Decaëns & Rossi, 2001), 而它们的生态位分异又过低, 所以种间竞争是这些蚯蚓空间格局的重要驱动过程(Jiménez et al, 2006; Decaëns et al, 2009)。物种共存分析被越来越多地用于检验种间竞争是否起调控作用(Gotelli & McCabe, 2002)。研究证明, 蚂蚁(Traniello & Levings,

1986; Gotelli & Ellison, 2002; Sanders et al, 2003)、蚯蚓(Fragoso & Rojas, 1997; Decaëns et al, 2008, 2009; Decaëns, 2010; Jiménez et al, 2012)、地表步甲(Brandl & Topp, 1985)、跳虫和螨类(Caruso et al, 2013)和陆地软体动物(Barker & Mayhill, 1999)的物种共同出现的实测值显著低于模拟值, 表明种间竞争是这些土壤动物群落维持的重要调控过程。但实际上群落内土壤动物食性复杂, 处于不同营养级。例如在德国多个森林生境内, 一个群落内的甲螨属于3–4个营养级(Schneider et al, 2004; Maraun et al, 2011), 用不考虑营养级关系的物种共存格局解释种间竞争作用时有一定局限。因此, 量化群落内物种的营养功能群是进一步验证土壤动物群落种间竞争作用的基础(Chang et al, 2016)。虽然许多土壤动物群落维持了很高的物种多样性, 但有关种内和种间竞争调控的证据非常缺乏(Wardle, 2006)。一方面, 土壤动物群落相同营养级之间相对较弱的竞争, 可能促使群落维持丰富的多样性(Wardle, 2006); 另一方面, 也说明种内和种间竞争对土壤动物群落的调控作用相当复杂, 还需要开展深入细致的研究。

捕食是土壤动物群落的显著调控过程。与线虫、线蚓、微型节肢动物和蚯蚓等次级消费者相比, 步甲、蜈蚣和蜘蛛等地表捕食者通常处于更高营养级(Goncharov & Tiunov, 2014), 它们不仅调控螨类和蚯蚓等较低营养级的消费者(Salmon et al, 2005; Gao et al, 2017b), 也调整局地营养级网络和维持群落内的碎屑功能单位(Goncharov & Tiunov, 2014)。捕食对土壤动物群落的调控表现复杂, 研究表明蜈蚣捕食显著地抑制跳虫数量(Salmon et al, 2005), 但也有研究表明蜈蚣并不影响跳虫数量(Gao et al, 2017b)。在许多生境特别是北方森林地表栖居的蜘蛛、隐翅虫和蜈蚣等捕食者有很高的丰富度(Petersen & Luxton, 1982), 然而有关这些捕食者对土壤动物群落调控作用的研究非常缺乏(Goncharov & Tiunov, 2014)。

资源竞争和天敌危害可能导致密度制约效应(density dependence)(祝燕等, 2009)。密度制约效应被认为是维持物种多样性的重要机制, 它对群落的调控作用已被公认(陈磊等, 2014)。密度制约假说强调群落内共存的物种具有生态位差异, 因而种内竞争远大于种间竞争(Levine & HilleRisLambers, 2009; 陈磊等, 2014); 而同种或亲缘关系较近的物种通过

竞争资源和化感作用等方式相互损害,从而为其他物种的生存提供了空间和资源,促进物种共存(王婷等, 2014)。密度制约调控不同土壤动物功能群的密度、数量和生长率等,包括地表捕食性蜘蛛(Takada & Miyashita, 2014)、蚂蚁及蚂蚁和蚜虫之间的共生关系(Addicott, 1979)、菌食性跳虫和捕食性螨类(Ferguson & Joly, 2010)、腐食性蚯蚓(Kammenga et al, 2003)等。密度制约还促进土壤动物集群及其巢穴空间格局的形成,当蚂蚁密度很高而能够利用的领域受限时,激烈的竞争可能会产生密度依赖空间格局(pattern of density-dependent spacing)(Ryti, 1991),但也有研究发现密度制约并不调控蚂蚁的空间格局(Schooley & Wiens, 2003)。密度制约还会调控捕食者-猎物之间的关系,如调控步甲幼虫的数量,导致步甲幼虫种内竞争的变化,进而改变其猎物鼻涕虫(*Deroceras reticulatum*和*Arion intermedius*)的密度(Thomas et al, 2009)。但密度制约对土壤动物群落的调控研究主要集中于少数类群,相关研究还有待于进一步开展。

除了种间竞争和捕食这两个研究较多的生物间作用(Hillebrand & Blenckner, 2002)外,非营养级生物间作用也起正的或者负的调控作用。土壤动物通过改变物理环境和创造特定微生境来间接调控其他土壤动物群落。如微型土壤节肢动物将凋落叶转换成粪粒,蚯蚓产生蚓粪,白蚁建造土堆,蚂蚁产生土堆巢穴,这些结构会产生异质性的有机质斑块或促进空间结构的变异性,从而影响其他土壤动物群落的空间分布和构建过程。蚯蚓可以为其他土壤动物营造适宜生境,蚯蚓数量多的生境中微型节肢动物的密度也随之增加(Hamilton & Sillman, 1989; Loranger et al, 1998)。因为蚯蚓活动增大土壤孔隙度,促进了其他土壤节肢动物扩散(Cameron et al, 2013),优化跳虫躲避捕食的逃跑路线,如蚯蚓粪便的聚集堆积为跳虫提供躲避捕食者的场所(Salmon et al, 2005),所以蚯蚓改善后的土壤孔隙度吸引了跳虫(Wickenbrock & Heisler, 1997),蚯蚓洞穴和蚯蚓粪增加了跳虫数量(Tiunov & Kuznetsova, 2000)和甲螨多样性(Loranger et al, 1998)。但是,不同营养级间的相互作用也会抑制其他土壤动物群落,如地下蚁巢降低食真菌线虫的密度(Blomqvist et al, 2000),蚯蚓活动降低微型螨类和跳虫的丰富度和物种多样性(Mclean & Parkinson, 2000)。可能

是因为蚯蚓取食了跳虫产在凋落物上的卵,从而抑制了其数量(Gao et al, 2017b),也可能是因为蚯蚓与跳虫、螨类的取食凋落物行为产生微弱的竞争(Brown, 1995),或者蚯蚓的取食和掘穴行为干扰了真菌的生长,降低了作为跳虫食物资源的真菌总量,从而间接抑制了跳虫数量(Chang et al, 2016; Gao et al, 2017b)。近年来级联效应(cascade effects),即捕食者多样性对食物网结构和动态的影响备受关注(邵元虎等, 2015)。蜘蛛、步甲等捕食者强烈影响其猎物的多度、多样性和行为(Miyashita & Niwa, 2006; Zhao et al, 2013),这些影响又通过猎物的取食逐步向更低营养级传递。这种自上而下的级联效应可以改变地下食物网内一个种群、群落或者整个营养级的丰富度、生物量和生产力(Pace et al, 1999),进而间接影响植物生物量、进入土壤生态系统的有机物质、养分循环特征等(Croll et al, 2005; Wardle et al, 2005; Miyashita & Niwa, 2006)。

3.2 环境过滤

土壤能提供三维的空间资源和高度异质的环境条件(Wardle, 2002),环境异质性增加微生境多样性和资源分异性,允许群落物种共存(Anderson, 1975; Giller, 1996; Schneider et al, 2004)。异质性环境比均质性环境能够维持更多的物种(Tamme et al, 2010),这种生态位的分化导致较高的土壤动物物种多样性,并将物种排序成相对异质的局地空间格局(Dornelas et al, 2006)。这种环境异质性导致土壤动物群落高密度和低密度斑块的镶嵌分布格局(Jiménez et al, 2006)。

土壤资源的空間异质性导致微生境的多样性,较大的资源分化促进物种共存。例如,在较小空间尺度的土壤和凋落物生境,小型节肢动物的物种丰富度随着微生境多样性的增加而增加(Anderson, 1978; Hansen & Coleman, 1998)。然而在土壤微宇宙实验中,小型土壤节肢动物多样性在4种资源充分混合的系统中很低,而在相同的4种资源构成的斑块系统中,其多样性却较高,这表明空间分化降低了竞争压力(OConnell & Bolger, 1998)。可见,土壤资源的空間异质性不仅能促进微生境多样性,也会使竞争性土壤动物发生空间分化。

小尺度空间环境条件相对均质,通常认为环境异质性对小尺度群落空间格局不起调控作用,但是单个植物和根系构型、动物的行为(Rossi, 2003a)和

微地形(Ettema & Wardle, 2002)都会强烈影响蚯蚓的空间格局。在40 cm × 60 cm的泥炭藓覆盖表面, 有壳变形虫在1–10 cm斑块内聚集, 几厘米范围内的微地形变异起到了重要作用(Mitchell et al, 2000)。即使地形和土壤环境相对均质, 由于该生境同时镶嵌在更大的空间尺度内, 土壤动物群落也在几厘米到几米的范围内表现出斑块状分布。如多年耕作和单一栽培种植对农田土壤产生同质性影响, 但主要线虫群落仍明显地聚集在6–80 m的空间斑块内(Robertson & Freckman, 1995), 说明人类耕种活动对胞囊线虫的空间格局有明显的影响(Webster & Boag, 1992)。

在局地尺度, 土壤动物群落常从十几米到上百米的范围内表现出空间格局(Ettema & Wardle, 2002)。首先, 这种空间格局受土壤理化性质、营养状态和污染状态空间异质性的影响(Ettema & Wardle, 2002; Bardgett, 2005; Khalil et al, 2009; Krab et al, 2010; Berg, 2012)。如农田土壤跳虫在大于200 m的范围内具有显著的空间格局, 这受较大尺度的土壤有机质和农业活动影响(Fromm et al, 1993)。美国乔治亚海岸平原的河岸湿地内, 一种食细菌线虫的分布斑块达到了67 m, 该空间格局受土壤湿度和硝态氮影响, 但这种影响强度随时间变化(Ettema et al, 1998); 而在苏格兰东部一个农业生态系统中, 线虫在5–50 m范围内的空间格局受人类农业活动的影响(Webster & Boag, 1992)。其次, 土壤动物群落的空间格局也受到植物异质性的调控, 包括植物大小、生长型和时间动态变化。如在混交林中, 不同树种的凋落物质量差异能够促进微生境异质性(Saetre, 1999; Saetrea & Bååthb, 2000), 因而地表植被的分异是土壤动物群落格局和构建的关键驱动力。另外, 土壤和植被等环境的异质性会同时调控土壤动物群落格局。美国科罗拉多州北部矮草草原的白蚁群落在330 m的范围内形成显著的空间格局, 这是由于较大尺度地形和植被空间异质性的影响(Crist, 1998)。擅长制造土堆的动物(如蚂蚁)会增加土壤环境异质性, 它们与植物交互作用进而影响土壤动物群落格局的斑块性(Ayarbe & Kieft, 2000; Blomqvist et al, 2000)。

在较大的景观尺度, 土壤湿度、pH值和资源可获得性的环境异质性强烈地影响温带森林土壤甲螨的群落组成和多样性(Maraun & Scheu, 2010)。但

也有研究发现在景观尺度, 营养状态、水的可获得性、土壤类型和植被类型的环境异质性对甲螨空间分布格局的解释能力微弱, 反而是地质年龄(Zaitsev et al, 2013)、气候和景观历史(Rantalainen et al, 2005; Zaitsev et al, 2006)对其分布格局有显著的调控作用。

3.3 随机扩散

大型土壤动物在土壤基质内的扩散能力相对较弱, 而中小型土壤动物可能通过其他(土壤)动物、水和风等媒介长距离被动扩散。土壤动物沿着土壤基质的主动扩散一般非常受限, 它们的每日移动距离从几毫米到几厘米, 不同种类的移动距离有所差异(Bengtsson, 1997; Dighton, 1997)。如线虫的移动受到土壤生境复杂性的影响(Jabbour & Barbercheck, 2008), 一些节肢动物在地下移动比在地上慢4倍(Kaczmarek, 1978)。当物种的移动距离比资源斑块小很多的时候, 许多物种能够在单一的资源环境中共存(Tilman & Kareiva, 1997), 土壤动物的多尺度空间格局便受到随机扩散过程的调控。即使在相对均质的生境, 繁殖和限制性扩散等内在过程也会导致土壤动物的聚集分布(Oconnell & Bolger, 1998)。如瑞士侏罗山高位沼泽内, 内在过程使有壳变形虫在13个厘米级尺度内形成聚集空间格局(Mitchell et al, 2000); 马提尼克(Martinique)一个草地生境中, 蚯蚓卵和成体扩散行为使其呈10–30 m的斑块状集群分布(Rossi et al, 1997)。

扩散对群落格局能产生不同的调控结果(Cadotte & Fukami, 2005), 因此很难解释其作用。不同斑块间的生境廊道可以促进不同生境间土壤动物的扩散, 在援救效应(Rescue effect)的调控下维持其他隔离生境的群落组成和格局。在微宇宙实验中(Rantalainen et al, 2004), 生境廊道显著增加了线蚓的多度和生物量, 但并没有显著提高土壤跳虫、螨类和线虫的多度, 可能是因为不同生境间的扩散距离与线蚓的扩散能力、体型大小相匹配, 而其他类群由于扩散能力较弱, 适应生存于扩散受限制的资源型斑块中。而野外调查实验发现跳虫扩散能力对群落维持有重要调控作用, 地表跳虫的强扩散能力使其能够快速重新定殖被洪水淹没过的生境, 使群落内部不同生境间的 β 多样性降低, 并增加群落内部物种组成的相似性(Russell & Griegel, 2006)。当空间范围扩大到区域尺度, 群落内不同生境间的隔

离度增加而连接度降低,随着空间距离增加,跳虫的扩散能力减弱,从而增加了不同生境之间物种组成的异质化,导致 β 多样性和 γ 多样性的增加(Starzomski & Srivastava, 2007)。

扩散能力和扩散模式是决定群落物种共存的关键。扩散能力强的物种因为更容易到达适宜生境,所以较少受到生态位和空间结构的限制(Padial et al, 2014),而扩散模式(主动扩散和被动扩散)对群落结构的影响较复杂(Bie et al, 2012)。土壤螨移动缓慢(Terauds et al, 2011),无论是主动扩散还是被动扩散的螨类,其空间格局均受扩散限制影响,因此螨类很难突破环境生态位的限制,其群落格局常受到环境异质性的调控(Mumladze et al, 2013)。具短翅的步甲飞翔能力相对较弱,群落内不同生境间 β 多样性差异较大,而具长翅的步甲移动能力较强,不同生境间 β 多样性差异较小(Hendrickx et al, 2009)。不同于土壤螨群落,地表步甲这种强扩散能力和主动扩散模式使其能够突破环境生态位的限制并到达适宜生境,从而削弱环境异质性对群落的调控。

长距离扩散是通过非常见扩散途径实现的低频率扩散事件,关注长距离扩散可促进对土壤动物群落分布格局和驱动机制的深入研究(Nathan, 2006)。土壤食根线虫可以通过风媒和水媒进行长距离扩散(Peña et al, 2011);土壤螨类和跳虫可以在海水表面和 underwater 存活 2 周左右,其被动扩散距离可达 700 km,理论上可以随着洋流从挪威北部扩散到斯瓦尔巴特群岛(Coulson et al, 2010)。这种强扩散能力和独特的繁殖模式使土壤动物广泛分布(Moore, 2002)。另外,地质历史事件、地理障碍和空间地理距离等也会影响土壤动物的随机扩散(Wiens & Milne, 1989),进而影响其空间格局和群落构建(Moore, 2002)。

农业、渔业和其他人为活动导致的外来物种入侵是一种不可忽视的随机过程,也对土壤动物群落构建起到重要的调控作用(Cameron et al, 2007; Wackett et al, 2018)。蚯蚓和蚂蚁是两类最主要的入侵土壤动物(Coyle et al, 2017),二者可以改变土壤结构、化学属性、土壤微生物群落并加剧土壤动物间资源竞争(Eisenhauer et al, 2007)。外来入侵蚯蚓可以显著降低土壤跳虫、螨等小型节肢动物的密度和物种数量(Migge, 2001; Eisenhauer et al, 2007),但可能通过改善微生境、提供避难所和提供食物

资源等增加线虫、革螨的丰富度(Maraun et al, 1999)。外来入侵蚂蚁通过直接捕食或资源竞争降低本地蚂蚁、跳虫和螨等土壤动物的丰富度(Walker, 2006),但在一些案例中,若时间充足,本地蚂蚁也会恢复到外来蚂蚁入侵之前的丰富度(Morrison, 2002)。外来入侵植物可能会改变进入本地土壤生态系统的植物资源的质量、数量、空间分布以及土壤温湿度和根系分布等,进而影响土壤线虫、节肢动物、蚯蚓等群落的物种丰富度和个体数量(Belnap et al, 2005; Belnap & Phillips, 2011; Coyle et al, 2017)。尽管已经开展了相关研究,但人类活动导致的外来物种入侵到底如何调控土壤动物多样性,其潜在的调控机制如何,仍然所知甚少(Wardle, 2006)。

我国土壤动物群落构建机制研究表明,土壤动物群落格局受到结构性因素、随机性因素或两者的共同影响,这说明土壤动物群落受到确定性和非确定性过程的共同调控。首先,研究表明随机扩散对土壤动物群落的调控尤其显著(高梅香等, 2014b; Sha et al, 2015),但这种扩散因类群而异:土壤螨较弱的扩散能力对群落构建有限制作用(Dong et al, 2017),而地表甲虫因有较强的扩散能力,群落在不同样点之间具有较高的物种周转率(高梅香等, 2018)。其次,虽然有研究表明环境过滤对群落组成没有很强的解释能力,但其调控作用仍不容忽视(张丽梅等, 2016)。土壤和地形等因子是影响土壤动物群落空间格局的重要因素,目前的研究可能忽略了某些更关键的环境过程,导致我们难以发现环境过滤的显著调控作用。第三,土壤动物群落表现为竞争型、聚集型和随机型多种格局(Lin et al, 2014; 高梅香等, 2016),种间竞争对土壤动物群落的调控作用并不普遍(Gao et al, 2017a; 倪娟平等, 2018a),但这并不能否认生物间作用的调控过程。

总之,目前研究表明生态位理论和中性理论是土壤动物群落多尺度格局和共存的重要机制(Gao et al, 2014),随机扩散和环境过滤的作用相对明显(Sha et al, 2015),而生物间作用的调控证据仍不充足(Gao et al, 2016)。

4 总结与展望

环境过滤、生物间作用和随机扩散对土壤动物群落构建的调控研究取得了很大进展,然而仍然面临一些挑战,如环境过滤的定义和应用受到一些学

者的批评(Kraft et al, 2015), 竞争、捕食和寄生等生物间作用对群落的调控能力研究主要集中于少数类群, 量化土壤动物群落的扩散过程仍然是一个难点。解决上述问题会极大地促进土壤动物群落构建机制的研究, 这也是未来发展的重要方向。

4.1 环境过滤对土壤动物群落的调控

经典的环境过滤被定义为环境条件选择那些有能力在给定地点存活和维持的物种(Emerson & Gillespie, 2008), 但这个概念忽视了生态学过程的复杂性。物种在某个群落的缺失反映的不仅是环境过滤导致的物种排除, 也有其他机制的作用(Kraft et al, 2015), 如扩散限制和局地竞争, 目前的研究常常不能区分这些机制(Cadotte & Tucker, 2017), 因此, 环境过滤的概念和应用应该被严格限定(Kraft et al, 2015)。受到这些观点的启发, 一些学者直接使用选择性过滤而不是环境过滤来表征生物因子和非生物因子对有机体适合度的影响(Graham & Stegen, 2017)。但也有研究认为在特定的背景下量化或应用环境过滤的概念仍然恰当(Thakur & Wright, 2017), 解释环境重要性时不应仅依赖于群落组成数据, 还应该综合考虑物种丰富度、种群生长或性状(Cadotte & Tucker, 2017)。另外, 根据系统发育距离或性状距离来分析“系统发育-环境”或“性状-环境”之间的关系, 是解释环境过滤是否起作用的有效途径。当群落物种具有相同的性状时, 可认为环境过滤起调控作用; 当群落内物种性状有明显变异时, 可认为生物间作用起调控作用。但生物间作用也可以导致与环境无关的性状产生相似性(Mayfield & Levine, 2010)。为了解释上述矛盾, Thakur和Wright (2017)提出一个框架, 即正确区别宏环境(macroenvironment)和微环境(microenvironment)之间的差别, 以及正确区分物种对宏环境的影响和这些影响反过来对物种的影响。关于环境过滤作用的深入研究, 会促进我们对土壤动物群落构建机制的认识(Kraft et al, 2015)。

4.2 基于食性关系的生物相互作用

虽然很多研究发现了生物间作用对土壤动物群落的调控作用, 但相关研究仍然非常缺乏。原因之一是很多研究将土壤生物群落视作一个黑箱或集中于特定的类群(Morrien et al, 2017), 如线蚓、蚯蚓、步甲和地表蜘蛛等, 而对于土壤动物群落整体网络和真正的营养级关系所知甚少。稳定同位素技

术能够确定不同时空尺度下食物来源、竞争关系、捕食-被捕食关系、食物网连通性、营养生态位宽度和摄食行为等(窦永静等, 2015; 陈展彦等, 2017)。明确群落内物种之间的基本营养关系是进一步阐明生物间作用对群落调控能力的基础。基于稳定同位素技术的分析是研究土壤动物群落食物网结构和营养级关系的有效途径(Tiunov, 2007), 会极大地促进生物间作用对土壤动物群落构建调控作用的研究(Gao et al, 2014)。

4.3 扩散对土壤动物群落构建的调控作用

量化土壤动物群落的扩散能力是一个难点, 也是土壤动物群落构建机制研究的一个障碍。土壤动物体型较小, 生存环境隐蔽性强, 无论是单个个体还是单个物种, 野外监测其扩散能力和扩散模式都极其困难(Heino et al, 2017)。一个群落常包含几十个物种, 在自然环境中直接获取群落内所有物种的扩散参数几乎是不可能的。借鉴相关研究成果, 土壤动物群落扩散能力可以通过多种指标来表征(Heino et al, 2017): (1)基于有机体的扩散指标, 包括土壤动物的体型大小、翅的有无、翅长或翅宽、扩散模式等。例如, 非常微小的土壤动物更容易被水、风或动物等搬运到较远的地方, 而体型较大的土壤动物被长距离搬运的可能性降低。翅长、翅宽是土壤动物扩散能力的指标, 如长翅步甲的扩散能力相对较强, 而短翅步甲的扩散能力相对较弱。以上指标的获得依赖于土壤动物功能性状参数, 即那些在个体水平可测量的, 通过影响物种的生长、繁殖或存活来影响物种适合度的生物特征(Violle et al, 2007), 反映了土壤动物对外部环境的适应和权衡。(2)基于分子遗传的扩散指标。种群遗传学、DNA条形码和环境DNA等生物技术和指标可以作为土壤动物及其群落的扩散指标。如不同谱系关系的地理分布能通过种群的空间遗传变异性来指示历史扩散事件。相关研究已证明基因结构和物种扩散能力有联系, 具有高基因多样性的种群的扩散能力相对较低(Heino et al, 2017)。(3)基于图形的扩散指标(Heino et al, 2017)。该方法可以评估地形结构和地理距离对土壤动物移动的影响, 如PCNM和Moran特征向量图等(Dray et al, 2006)。

土壤动物是研究群落生态学基本问题的理想实验对象(Gerisch, 2011), 但是以土壤动物为对象的群落构建理论的验证非常少。群落构建机制假说

或理论包括Lotka-Volterra竞争模型、竞争-拓殖权衡假说、资源比例假说、更新生态位假说、中性理论、近中性理论(牛克昌等, 2009)和集合群落理论(Leibold et al, 2004)等, 目前以土壤动物为对象的群落构建机制研究, 多基于模型验证生态位理论和中性理论、集合群落理论、确定性过程和非确定性过程的调控作用, 而对其他理论的深入解释和验证明显不足。土壤动物群落格局和构建机制研究已取得重要进展, 我国相关研究虽然刚刚起步, 但有很大的发展潜力, 需要土壤动物学家共同努力, 促进该领域的深入发展。

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